

Information on measures and related costs in relation to species considered for inclusion on the Union list: *Prosopis juliflora*

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Species (scientific name)	<i>Prosopis juliflora</i> (SW.) DC
Species (common name)	Mesquite, algarroba
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Date Completed	01 October 2018
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Summary

Highlight of measures that provide the most cost-effective options to prevent the introduction, achieve early detection, rapidly eradicate and manage the species, including significant gaps in information or knowledge to identify cost-effective measures.

Prosopis juliflora is a legume tree or shrub native to northern South America, Central America and the Caribbean. It was introduced over the past two centuries, mostly to tropical drylands in Africa, Asia and Oceania, and widely planted in reforestation schemes, especially in the 1980s and 1990s, by international organizations and national authorities; these latter introductions became the source of many of the largest invasions. *Prosopis juliflora* is by far the dominant invasive species in the genera, especially in tropical regions, though it is occasionally found alongside the much less invasive *P. pallida* in some restricted areas. In more sub-tropical regions (e.g. southern Africa and Australia, as well as in their native North American range), other species dominate (notably *Prosopis glandulosa*, *Prosopis velutina* and hybrids) and are also highly invasive.

In 2016, *P. juliflora* was prioritized (amongst 36 species from the EPPO List of Invasive Alien Plants and a horizon scanning study) for PRA within the LIFE funded project “Mitigating the threat of invasive alien plants to the EU through pest risk analysis to support the Regulation 1143/2014’ (see www.iap-risk.eu). It was also one of 16 species identified as having a high priority for PRA. The species is certainly one of the most invasive woody weeds in the world’s tropical drylands, and the genus as a whole was included in the widely cited ‘100 of the World’s Worst Invasive Alien Species’. In a review of introductions of *Prosopis* species globally, Shackleton, Le Maitre, van Wilgen and Richardson (2014) found that 79% of introductions led to naturalization, of which 38% then became invasive.

There are only very few reports of any *Prosopis* species naturalizing in European Union countries. For *P. juliflora*, the only known reports for presence are of two planted trees in a sheltered valley in Almeria, south-eastern Spain (Pasiiecznik and Peñalvo López, 2016), and reported as naturalised in a very limited area in Gran Canaria in the Canary Islands, Spain (Verloove, 2013, 2017). This author considers that the likelihood of other *Prosopis* species (e.g. *P. chilensis*, *P. glandulosa* and *P. velutina*) becoming invasive in Union countries is significantly higher than for *P. juliflora*. However, all the above species are very closely related, and proposed measures for management and control for *P. juliflora* contained herein would therefore also be relevant for any of the aforementioned species.

Based on current environmental conditions and species distribution modeling developed and used in the recent PRA for the EPPO region (EPPO, 2018), a number of suitable areas for establishment of *P. juliflora* were identified. This particularly includes the Mediterranean and Macaronesian biogeographical region of the Union, in largely frost-free coastal and low-lying inland areas. This includes parts of Cyprus, Greece (and the islands), Italy (including Sardinia and Sicily), Malta, Portugal (including Madeira and the Azores), and Spain (including the Canary Islands). Results of the PRA also concluded that *P. juliflora* poses a moderate risk to the endangered area (Mediterranean and Macaronesian biogeographical region) with a moderate uncertainty (EPPO, 2018). The major pathway to be considered is **Plants for planting**, for use in reforestation and as an ornamental. Given the significant impact of this (and closely related) species in other parts of the world and the identified risk to Union countries, a number of management needs should be considered. In summary, this requires regulation, and suggested measures are detailed in subsequent sections. Also, national measures should be combined with international measures, and international coordination of management of the species between countries is recommended.

- (1) Prevention of intentional introduction and spread – the prohibition of import, sale and movement of plants and seeds, as would be required under Article 7 of the EU IAS Regulation 1143, if the species were to be listed.
- (2) Prevention of unintentional introduction and spread – not applicable.
- (3) Prevention of secondary spread of the species – Removal of naturalized individuals and populations where known to exist as prevention of secondary spread once well-established over a large area is not possible.
- (4) Surveillance measures to support early detection – Undertaking full surveys in the endangered area, including full literature reviews, with an obligation to report findings if the species was regulated.
- (5) Rapid eradication of new introductions – manual eradication to remove all identified plants.
- (6) Management – manual control, and where widespread, countries must prepare and implement eradication and containment/management plans (that could also include mechanical, chemical and/or biological control methods).

Prevention of intentional introductions and spread – measures for preventing the species being introduced intentionally. **This table is repeated for each of the prevention measures identified.** *If the species is listed as an invasive alien species of Union concern, this table is not needed, as the measure applies anyway.*

Measure description
Provide a description of the measure, and identify its objective

The prohibition of import, sale and movement of plants and seed, as would be required under Article 7 of the EU IAS Regulation 1143, if the species were to be listed.

The major pathway of introduction for the species to be addressed is **Plants for planting**. Therefore, to prevent intentional introductions of the species into the EU, the prohibition of import into and movement into the EU would be required, alongside banning the sale and planting of plants and seed of plants labeled or otherwise identified as *P. juliflora*.

<p>Effectiveness of the measure Is it effective in relation to its objective? Has the measure previously worked, failed?</p> <p>Please select one of the categories of effectiveness (with an 'X'), and provide a rationale, with supporting evidence and examples if possible.</p>	<table border="1"> <tr> <td><i>Effectiveness of measures</i></td> <td><i>Effective</i></td> <td>X</td> <td><i>Neutral</i></td> <td></td> <td><i>Ineffective</i></td> </tr> </table>						<i>Effectiveness of measures</i>	<i>Effective</i>	X	<i>Neutral</i>		<i>Ineffective</i>
	<i>Effectiveness of measures</i>	<i>Effective</i>	X	<i>Neutral</i>		<i>Ineffective</i>						
<p><i>Rationale:</i> Prevention of intentional introduction is the only fully effective measure, as once the species is established/naturalized over more than a limited area, eradication is considered impossible (Pasiecznik et al., 2001). <i>Prosopis</i> species are also on the regulated list of other countries (see below). And, as to date, no alien <i>Prosopis</i> species are reported as naturalized on the continental USA and no further <i>Prosopis</i> species reported in Australia, these measures have proved effective.</p> <p>Australia - <i>Prosopis</i> spp. (as a genus) is listed as one of the 30 Weeds of National Significance (www.environment.gov.au/cgi-bin/biodiversity/invasive/weeds/weeddetails.pl?taxon_id=68407), and includes <i>P. juliflora</i> as one of four naturalized species (the others being <i>P. glandulosa</i>, <i>P. pallida</i> and <i>P. velutina</i>, and hybrids).</p> <p>South Africa - <i>Prosopis juliflora</i> is not listed as invasive, but under the country's National Environmental Management and Biodiversity Act (NEMBA), <i>P. glandulosa</i>, <i>P. velutina</i> and their hybrids are listed as Category 1b (may not be owned, imported or grown) in Eastern Cape, Free State, North-West and Western Cape, and Category 3 (may hold but cannot propagate or sell) in Northern Cape (www.environment.co.za/weeds-invaders-alien-vegetation/alien-invasive-plants-list-for-south-africa.html#notice1).</p> <p>USA - <i>Prosopis juliflora</i> is not included in the USDA Federal noxious weed list (last updated 21 March 2017, www.aphis.usda.gov/plant_health/plant_pest_info/weeds/downloads/weedlist.pdf), although 20 of the 44 <i>Prosopis</i> species recognized by Burkart (1976) are listed, 16 as A1 weeds and 4 as A2 weeds. The reasons for not being included is unclear, however, but may be due, perhaps, to the mistaken view that <i>P. juliflora</i> is native to the USA, following Bentham's classification. Other native species (<i>P. glandulosa</i> and <i>P. velutina</i>) are also not listed. However, one US state,</p>												

Hawaii, does include *P. juliflora* on its list of noxious weeds (see, Division of Plant Industry. List of plant species designated as noxious weeds (20 October 2003). Hawaii Department of Agriculture, Hawaii. (in <https://plants.usda.gov/java/reference?symbol=PRJU3>)). Many other states contain the same species as listed in the federal USDA, with some variation, e.g. the California State-listed noxious weeds (<https://plants.usda.gov/java/noxious?rptType=State&statefips=06>) includes *P. velutina* as the preferred name for *P. articulata* (whereas Burkart (1976) considered them as separate species and not synonyms). The whole genus is listed as a noxious weed in the State of Florida (<https://plants.usda.gov/java/noxious>).

Detection and identification

The following description is taken from Burkart (1976) as the over-arching species morphology, including all varieties from all parts of the world. Although some material that Burkart (1976) identified as *P. juliflora* is now likely to be *P. pallida* (Harris, Pasiecznik, Smith, Billington & Ramirez, 2003), this description is still accepted in the absence of a newer acknowledged taxonomy.

Prosopis juliflora is a tree 3-12 m tall, sometimes shrubby with spreading branches; wood hard; branches cylindrical, green, more or less round- or flat-topped, somewhat spiny with persistent, green (sometimes glaucous or greyish, not reddish) foliage, glabrous or somewhat pubescent or ciliate on the leaflets; spines axillary, uninodal, divergent, paired, or solitary and paired on the same branches, sometimes absent, not on all branchlets, measuring 0.5-5.0 cm long, being largest on strong, basal shoots. Leaves bipinnate, glabrous or pubescent, 1-3 pairs of pinnae, rarely 4 pairs; petiole plus rachis (when present) 0.5-7.5 cm long; pinnae 3-11 cm long; leaflets 6 to 29, generally 11 to 15 pairs per pinna, elliptic-oblong, glabrous or ciliate, rarely pubescent, approximate on the rachis or distant a little more than their own width, herbaceous to submembranous (not sub-coriaceous as in more xerophilous species and therefore often corrugated or curved when dried), emarginated or obtuse, pinnate-reticulately curved; leaflets 6-23 mm long x 1.6-5.5 mm wide. Racemes cylindrical, 7-15 cm long, rachis puberulent; florets as usual, greenish-white, turning light yellow. Legume straight with incurved apex, sometimes falcate, straw-yellow to brown, compressed, linear with parallel margins, stalked and acuminate, 8-29 cm long x 9-17 mm broad x 4-8 mm thick; stipe to 2 cm; endocarp segments up to 25, rectangular to subquadrate, mostly broader than long; seeds oval, brown, transverse.

Prosopis species, however, exhibit high levels of variability in morphological characters in its native range. Self-incompatibility and obligate outcrossing tend to lead to large phenological variation, as a combination of both clinal variation in response to broad climatic factors and ecotypic (discontinuous) variation in response to disjunct environmental factors. Differences in continuous climatic clines such as temperature, rainfall and day length, and discrete differences in site such as soil type, salinity or depth combine to create a variety of phenological responses.

Identifying Tropical *Prosopis* Species: A field guide (Pasiiecznik, Harris & Smith 2004) provides the easiest to use means of separating the eight most common *Prosopis* species found in tropical regions, from field observations and measurements of morphological characteristics. It also includes a description of the most common misidentifications, and a simple key to separate *P. juliflora* and *P. pallida* using leaf/leaflet size and number. In addition, the fact that *P. juliflora* is confirmed as the only tetraploid species in the genus means that flow cytometry analyses of genome size can be used as a tool from separating this species from others (Trenchard, Harris, Smith & Pasiiecznik, 2008).

However, ongoing taxonomic confusion surrounding *Prosopis* species within Section Algarobia must be highlighted, as this would impact on any proposed regulation, and some databases group all *Prosopis* species together or repeat taxonomical errors of the past. Furthermore, the general common name is mesquite or simply prosopis. Note also that as a common name, species of *Prosopis* are also referred to in normal script (not italics) and all in lower case, as are acacia, eucalyptus, leucaena, etc. In addition, as a common name, mesquite is also used for other species of Section Algarobia such as *P. glandulosa* (Lowe, Browne, Boudjelas & De Poorter, 2000), and occasionally for others outside of this Section, either with or without a specific epithet (e.g. *P. glandulosa* should be honey mesquite, *P. velutina*, velvet mesquite, etc.).

The following information on taxonomy and nomenclature is adapted from the *P. juliflora* datasheet in the Invasive Species Compendium (CABI, 2018; prepared by this author), the most up-to-date review of the taxonomy of species. *Prosopis juliflora* (Sw.) DC. has had an array of synonymy since its first description in 1788. Originally known as *Mimosa juliflora* Sw., it became both *Algarobia juliflora* (Sw.) Benth. ex Heynh. and *Neltuma juliflora* (Sw.) Raf. during the last two centuries before both genera were incorporated into the single, overarching genus *Prosopis*. Bentham (1875) noted *P. limensis* (syn. *P. pallida*) from Peru as the only *Prosopis* species of section Algarobia he was aware of that was not sympatric with others in the section. This may assume that he was either unaware of *P. juliflora* and hybrids in Ecuador and northern Peru, or that he treated them all as the same species, distinct from the *P. juliflora* of Central America, Colombia and the Caribbean.

Prosopis juliflora was used by Pasiiecznik et al. (2001) in its original, restricted and certainly biological sense, re-established by Burkart (1940) and accepted by Benson (1941) and Johnston (1962). The all-embracing, collective *P. juliflora* concept of Bentham (1875) was maintained by others and though this is rejected by most taxonomists, it is still used occasionally to this day. Confusion also occurs when referring to old literature, as the binomial *P. juliflora* was used to describe species now generally accepted as separate taxa. The following three varieties were accepted by Burkart (1976) and without any information to the contrary, also by Pasiiecznik et al. (2001): *Prosopis juliflora* (Sw.) DC. var. *juliflora*, *Prosopis juliflora* (Sw.) DC. var. *inermis* (H.B.K.) Burkart and *Prosopis juliflora* (Sw.) DC. var. *horrida* (Kunth) Burkart. However, even then, the taxonomy seemed uncertain, with Burkart noting that var. *inermis* and var.

	<p><i>horrida</i>, differed from var. <i>juliflora</i> principally in the relative presence/absence of thorns, with no other striking morphological basis for separation. However, particularly at the limits of the native range, further revision is expected.</p> <p>The '<i>P. pallida</i> – <i>P. juliflora</i> complex' was proposed by Pasiecznik et al. (2001) as a means to overcome the observed ambiguities at that time and the lack of agreement on how to deal taxonomically with tropical American <i>prosopis</i>, and discusses previous proposals and revisions in detail. This followed the treatment by Johnston (1962), who divided <i>P. juliflora</i> into two races, the Central American, and Colombian-Caribbean race, mainly on the basis of leaflet length, and noted the similarities and the differences between these two and the truly South American <i>P. limensis</i> (syn. <i>P. pallida</i>). However, it has since been unequivocally shown that the two are distinct taxa, morphologically and genetically (e.g. Harris et al., 2003; Landeras, Alfonso, Pasiecznik, Harris & Ramirez, 2006; Catalano, Vilaridi, Tosto & Saidman, 2008; Trenchard et al., 2008; Palacios et al. 2012; Sherry et al., 2012). Comparing native range material with that from introduced populations, however, highlighted a number of serious misidentifications, notably being that the 'common' <i>prosopis</i> in the north east of Brazil, Cape Verde and parts of Senegal is in fact <i>P. pallida</i>, and not <i>P. juliflora</i> as it has always been referred to (Harris et al., 2003). <i>P. pallida</i> has also been positively identified in southern Mauritania (Pasiecznik et al., 2006) and Djibouti (Pasiecznik, 2013) from naturalized populations. However, notwithstanding this published literature, scientific publications from Brazil and Cape Verde, for example, still tend to incorrectly refer to <i>P. juliflora</i> as the dominant species there (e.g. Fonseca, Albuquerque, Leite & Lira, 2016; Tavares & Barros, 2016).</p>						
<p>Side effects (incl. potential) – both positive and negative i.e. positive or negative side effects of the measure on public health, environment including non-targeted species, etc. For each of the side effect types please select one of the impact categories (with an 'X'), and provide a rationale, with supporting evidence and examples if possible.</p>	<p>Environmental effects</p>	<p>Positive</p>		<p>Neutral or mixed</p>	<p>X</p>	<p>Negative</p>	
	<p>Social effects</p>	<p>Positive</p>		<p>Neutral or mixed</p>	<p>X</p>	<p>Negative</p>	
	<p>Economic effects</p>	<p>Positive</p>		<p>Neutral or mixed</p>	<p>X</p>	<p>Negative</p>	
	<p>Rationale: There are no known environmental, social and economic side effects expected from the implementation of these measures, involving only the additional of a further species [or group of species] to the list of those plants that are regulated in the EU, and associated checks.</p>						
<p>Acceptability to stakeholders e.g. impacted economic activities, animal welfare considerations, public perception, etc. Please select one of the categories of acceptability (with an 'X'), and</p>	<p>Acceptability to stakeholders</p>	<p>Acceptable</p>		<p>Neutral or mixed</p>	<p>X</p>	<p>Unacceptable</p>	
	<p>Rationale: <i>Prosopis juliflora</i> and other species are known to have benefits and costs in other regions of the world where they are invasive. This has led to contentious issues between stakeholders. However, <i>Prosopis</i> species are not widely planted in the EU, but as it is known as a street tree in other countries, that other <i>Prosopis</i> species are planted as</p>						

<p>provide a rationale, with supporting evidence and examples if possible.</p>	<p>ornamentals, and that <i>Prosopis</i> seed and plants are sold by commercial companies, there may be some resistance to regulation from commercial suppliers. But as it is only a very minor ornamental species, this is not considered as a significant factor. As such, as <i>Prosopis</i> species are not planted in the EU for ornamental or grown for any other reason, no objection may be expected from commercial suppliers or the public.</p>						
<p>Additional cost information¹ When not already included above, or in the species Risk Assessment. - the cost of inaction - the cost-effectiveness - the socio-economic aspects Include quantitative &/or qualitative data, and case studies (incl. from countries outside the EU).</p>	<p>Limited information is available on quantitative costs for action or inaction, though some references exist, such as in Ethiopia (Wakie, Hoag, Evangelista, Luizza & Laituri, 2016) and South Africa (Wise et al., 2012). However, a recent PRA (EPPO, 2018) reported that impacts would be restricted to only small areas in the EU where <i>P. juliflora</i> can establish, but that in the absence of specific data on impacts the rating of magnitude “remains high for impacts on biodiversity, ecosystem services and socio-economic impacts, however, uncertainty is raised too high for all categories, as it is not clear if these impacts will be realised throughout areas of potential establishment...” In addition, it notes that “In the EU, in frost-free coastal and low-lying inland areas of Cyprus, Greece, Italy, Malta, Portugal, and Spain, impacts on biodiversity and impacts on ecosystem services could be similar to those impacts seen in the current area of distribution and the isolated areas of establishment in the EPPO region, with the exception, potentially, of significant impacts on communities and local livelihoods... However, for this to be realised extensive populations of the species would need to establish and this would be more uncertain of occurring compared to areas in Israel and Jordan. In addition, even though the species has been sold as an ornamental species and as a forestry species globally, this is unlikely to be a significant pathway into the EU in future. Therefore, for EU Member States detailed in the endangered area (as above) a moderate rating has been given for impacts on biodiversity, ecosystem services and socio-economic impacts with a high uncertainty.” The PRA (EPPO, 2018) concluded that “the risk of introduction [of <i>P. juliflora</i>] and the potential area for establishment are both perceived as low, leading the EWG [expert working group brought together by EPPO to undertake the PRA] to propose an overall phytosanitary risk score of moderate.”</p>						
<p>Level of confidence on the information provided² Please select one of the confidence categories along with a statement to support the category chosen. See <i>Notes</i> section at the bottom of this document. NOTE – this is not related to the effectiveness of the measure</p>	<p><i>Inconclusive</i></p>	<p><i>Unresolved</i></p>	<p><i>Established but incomplete</i></p>	<p><i>Well established</i></p>	<p>X</p>	<p><i>Rationale:</i> High confidence is based in the knowledge obtained from a number of previous PRAs, including the following. <i>P. juliflora</i></p> <ul style="list-style-type: none"> Express Pest Risk Analysis for <i>P. juliflora</i> for the EU/EPPO region (2018) undertaken for the LIFE project (awaiting final validation). Australian/New Zealand Weed Risk Assessment adapted for Hawai’i (2005), - High risk, Score 19. www.hear.org/pier/wra/pacific/prosopis_juliflora_htmlwra.htm. <p><i>P. glandulosa</i></p> <ul style="list-style-type: none"> Spain –Score 22 and 32, ranking 6th and 4th in a list of 80 potential invasive plants, assessed by WRA and WG-WRA, respectively (Andreu & Vilà, 2010). 	

	<ul style="list-style-type: none"> Hawaii/Pacific - High risk, Score 19 (www.hear.org/pier/wra/pacific/Prosopis%20glandulosa.pdf) <i>P. spp.</i> Australia - Reject, Score 20 (www.hear.org/pier/wra/australia/prosp-wra.htm) <p>In addition, a detailed datasheet can be found in CABI's Invasive Species Compendium (https://www.cabi.org/isc/datasheet/43942).</p>
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Prevention of un-intentional introductions and spread – measures for preventing the species being introduced un-intentionally (cf. Article 13 of the IAS Regulation). **This table is repeated for each of the prevention measures identified.**

Measure description Provide a description of the measure, and identify its objective	Not applicable. Unintentional introduction of <i>Prosopis</i> seed as a contaminant is considered very unlikely. The only other possibilities for unintentional introduction are via live livestock imports where the animals have been fed on <i>Prosopis</i> pods either just before, or during, transit. Oceanic dispersal into the EU is also a possibility, but the risk is considered very low.				
Scale of application At what scale is the measure applied? What is the largest scale at which it has been successfully used? Please provide examples, with areas (km ² or ha) if possible.	n/a				
Effectiveness of the measure Is it effective in relation to its objective? Has the measure previously worked, failed? Please select one of the categories of effectiveness (with an 'X'), and provide a rationale, with supporting evidence and examples if possible.	Effectiveness measures	of	Effective	Neutral	Ineffective
	Rationale: n/a				
Effort required e.g. period of time over which measure need to be applied to have results	n/a				
Resources required ¹ e.g. cost, staff, equipment etc.	n/a				

Side effects (incl. potential) – both positive and negative i.e. positive or negative side effects of the measure on public health, environment including non-targeted species, etc. For each of the side effect types please select one of the impact categories (with an 'X'), and provide a rationale, with supporting evidence and examples if possible.	Environmental effects	<i>Positive</i>	<input type="checkbox"/>	<i>Neutral or mixed</i>	<input type="checkbox"/>	<i>Negative</i>	<input type="checkbox"/>	
	Social effects	<i>Positive</i>	<input type="checkbox"/>	<i>Neutral or mixed</i>	<input type="checkbox"/>	<i>Negative</i>	<input type="checkbox"/>	
	Economic effects	<i>Positive</i>	<input type="checkbox"/>	<i>Neutral or mixed</i>	<input type="checkbox"/>	<i>Negative</i>	<input type="checkbox"/>	
	<i>Rationale:</i> n/a							
Acceptability to stakeholders e.g. impacted economic activities, animal welfare considerations, public perception, etc. Please select one of the categories of acceptability (with an 'X'), and provide a rationale, with supporting evidence and examples if possible.	Acceptability to stakeholders	<i>Acceptable</i>	<input type="checkbox"/>	<i>Neutral or mixed</i>	<input type="checkbox"/>	<i>Unacceptable</i>	<input type="checkbox"/>	
	<i>Rationale:</i> n/a							
Additional cost information ¹ When not already included above, or in the species Risk Assessment. - implementation cost for Member States - the cost of inaction - the cost-effectiveness - the socio-economic aspects Include quantitative &/or qualitative data, and case studies (incl. from countries outside the EU).								
Level of confidence on the information provided ² Please select one of the confidence categories along with a statement to support the category chosen. See <i>Notes</i> section at the bottom of this document.	<i>Inconclusive</i>	<input type="checkbox"/>	<i>Unresolved</i>	<input type="checkbox"/>	<i>Established but incomplete</i>	<input type="checkbox"/>	<i>Well established</i>	<input type="checkbox"/>
	<i>Rationale:</i>							

NOTE – this is not related to the effectiveness of the measure

Prevention of secondary spread of the species – measures for preventing the species spreading once they have been introduced (cf. Article 13 of the IAS Regulation). This table is repeated for each of the prevention measures identified.

Measure description

Provide a description of the measure, and identify its objective

The prevention of secondary spread, once the species is well established, is deemed to be not possible. The only option is the removal of naturalized individuals and populations where known to exist [The specific measures to achieve this objective are described in the Rapid eradication and Management tables below.]

Once established over a large area, it has been shown that prevention of further spread of *P. juliflora* is not possible as the species quickly builds up a considerable seed bank, requiring regular removal of all new seedlings over very many years, as seeds can remain viable for at least 40 years and probably much longer (Pasicznik et al., 2001), and seeds spread easily by water and animals, with rates of spread in South Africa noted at around 14% per annum (Wise et al., 2012).

Means of spread described below cannot be realistically or effectively reduced. The only possible way would be to fence off effective areas thus prevent entry of livestock and large wild animals, but smaller mammals could still cause spread and thus is deemed relatively ineffective.

Natural (non-biotic) dispersal - Water is an important dispersal agent in desert ecosystems. Water dispersal ensures widespread dissemination of seeds during flooding or other high rainfall events when seedling establishment is favoured. *Prosopis* species are often found colonizing ephemeral watercourses and dispersal is aided by water flow in the rainy season, particularly during very wet years (Solbrig & Cantino, 1975). Oceanic dispersal is also important in coastal areas, and possibly for crossing large bodies of water such as in the Caribbean. Pods and endocarps float and are impervious to water infiltration, protecting the seed from the harmful effects of extended periods in sea water.

Vector transmission (biotic) - Pods have a high sugar content, are low in anti-feedants, and are widely sought after by a variety of animals. Disjunct stands of trees near to old centres of population suggest that man has also been a dispersal agent in historic and prehistoric times. Livestock are now the primary dispersal agents, although the pods are also avidly consumed by a wide variety of wild animals that play a major role in seed dispersal. Birds, bats, reptiles and ants also feed on *Prosopis* fruits and are potential, if only minor, agents of dispersal, but it is generally accepted

	<p>that the fruits and seeds are specialized for animal dispersion. Pods are eaten off the tree or off the ground and seeds are deposited in the faeces. Voided seeds are given a positive advantage by being placed in faeces, with their improved water-holding capacity and high levels of nutrients. Livestock may tend to spend more time on better pasture or by water sources but voiding of seeds in preferential locations is not guaranteed. However, different animals have very different effects on seed survival.</p> <p>Agricultural practices - Pods and seeds may adhere to agricultural machinery, but this is considered as a minimal cause of spread. The principal reason for agriculture increasing spread is due to habitat modification (e.g. resulting from overgrazing), which creates favourable conditions for the spread of <i>Prosopis</i>.</p>																					
<p>Scale of application At what scale is the measure applied? What is the largest scale at which it has been successfully used? Please provide examples, with areas (km² or ha) if possible.</p>	<p>The measure should be applied in areas where <i>P. juliflora</i> is known to be present in the EU, i.e. currently, only in a small area of Gran Canaria (Canary Islands) and Almeria, Spain.</p>																					
<p>Effectiveness of the measure Is it effective in relation to its objective? Has the measure previously worked, failed? Please select one of the categories of effectiveness (with an 'X'), and provide a rationale, with supporting evidence and examples if possible.</p>	<table border="1" data-bbox="645 683 1888 746"> <tr> <td>Effectiveness of measures</td> <td>Effective</td> <td></td> <td>Neutral</td> <td></td> <td>Ineffective</td> <td>X</td> </tr> </table> <p><i>Rationale:</i> See Rapid eradication and Management tables below.</p>	Effectiveness of measures	Effective		Neutral		Ineffective	X														
Effectiveness of measures	Effective		Neutral		Ineffective	X																
<p>Effort required e.g. period of time over which measure need to be applied</p>	<p>See Rapid eradication and Management tables below.</p>																					
<p>Resources required¹ e.g. cost, staff, equipment etc.</p>	<p>See Rapid eradication and Management tables below.</p>																					
<p>Side effects (incl. potential) – both positive and negative i.e. positive or negative side effects of the measure on public health, environment including non-targeted species, etc.</p>	<table border="1" data-bbox="645 1161 1888 1257"> <tr> <td>Environmental effects</td> <td>Positive</td> <td>X</td> <td>Neutral or mixed</td> <td></td> <td>Negative</td> <td></td> </tr> <tr> <td>Social effects</td> <td>Positive</td> <td>X</td> <td>Neutral or mixed</td> <td></td> <td>Negative</td> <td></td> </tr> <tr> <td>Economic effects</td> <td>Positive</td> <td>X</td> <td>Neutral or mixed</td> <td></td> <td>Negative</td> <td></td> </tr> </table> <p><i>Rationale:</i> See Rapid eradication and Management tables below.</p>	Environmental effects	Positive	X	Neutral or mixed		Negative		Social effects	Positive	X	Neutral or mixed		Negative		Economic effects	Positive	X	Neutral or mixed		Negative	
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<p>For each of the side effect types please select one of the impact categories (with an 'X'), and provide a rationale, with supporting evidence and examples if possible.</p>										
<p>Acceptability to stakeholders e.g. impacted economic activities, animal welfare considerations, public perception, etc. Please select one of the categories of acceptability (with an 'X'), and provide a rationale, with supporting evidence and examples if possible.</p>	<p>Acceptability to stakeholders</p>	<p><i>Acceptable</i></p>	<p>X</p>	<p><i>Neutral or mixed</i></p>		<p><i>Unacceptable</i></p>		<p><i>Rationale:</i> See Rapid eradication and Management tables below.</p>		
<p>Additional cost information¹ When not already included above, or in the species Risk Assessment. - implementation costs - the cost of inaction - the cost-effectiveness - the socio-economic aspects Include quantitative &/or qualitative data, and case studies (incl. from countries outside the EU).</p>	<p>None available. See Rapid eradication and Management tables below.</p>									
<p>Level of confidence on the information provided² Please select one of the confidence categories along with a statement to support the category chosen. See <i>Notes</i> section at the bottom of this document.</p>	<p><i>Inconclusive</i></p>		<p><i>Unresolved</i></p>		<p><i>Established but incomplete</i></p>	<p>X</p>	<p><i>Well established</i></p>		<p><i>Rationale:</i> See Rapid eradication and Management tables below.</p>	

Surveillance measures to support early detection - Measures to run an effective surveillance system for achieving an early detection of a new occurrence (cf. Article 16). This section assumes that the species is not currently present in a Member State, or part of a Member State's territory. **This table is repeated for each of the early detection measures identified.**

<p>Measure description Provide a description of the measure, and identify its objective</p>	<p>Undertaking full active surveys in the endangered area, following a review of literature to identify high risk areas, with an obligation to report findings if the species was regulated. <i>Prosopis juliflora</i> should be placed on NPPO's alert lists and they should report any findings, with increased surveillance in areas where there is a high risk the species may invade. NPPO's should also provide land managers and stakeholders with identification guides and information on control techniques and management, and facilitate regional cooperation.</p> <p>All known populations, and individuals when presence is very restricted, should have their GPS coordinates recorded. Remote sensing has also proved effective in other countries to assess the scale of invasion and identify potential areas of invasion (e.g. Maroni et al., 2016), but results require ground-truthing over the areas where invasion is known. It is not believed to be an effective measure for early detection (and therefore a separate table is not provided). Remote sensing using satellite imagery data to map distribution of <i>P. juliflora</i> has been used in a number of countries, e.g. Ethiopia (Wakie et al., 2016), Somalia (Maroni et al., 2016). This method can be applied to relatively large areas, e.g. by Wakie et al. (2016) to 95,266 km² in the Afar Region of Ethiopia, and by Maroni et al. (2016) to 5,167 km² in Somalia. However, it is not considered that remote sensing would be required over the restricted areas where <i>P. juliflora</i> is found in the EU.</p>					
<p>Scale of application At what scale is the measure applied? What is the largest scale at which it has been successfully used? Please provide examples, with areas (km² or ha) if possible.</p>	<p>This measure should be applied in all countries in endangered areas (including parts of Cyprus, Greece, Italy, Malta, Portugal, Spain,), in areas identified following a full review of literature and targeted interviews, with obligations to report finding as stipulated with the NPPOs of these countries.</p>					
<p>Effectiveness of the measure Is it effective in relation to its objective? Has the measure previously worked, failed? Please select one of the categories of effectiveness (with an 'X'), and provide a rationale, with supporting evidence and examples if possible.</p>	<table border="1" data-bbox="629 1078 1890 1150"> <tr> <td data-bbox="629 1078 927 1150"><i>Effectiveness of measures</i></td> <td data-bbox="927 1078 1151 1150"><i>Effective</i></td> <td data-bbox="1151 1078 1211 1150">X</td> <td data-bbox="1211 1078 1509 1150"><i>Neutral</i></td> <td data-bbox="1509 1078 1830 1150"><i>Ineffective</i></td> </tr> </table> <p><i>Rationale:</i> NPPOs are accustomed to implementing obligations to report findings thus should be effective, but effectiveness cannot be ascertained. However, noting the restricted distribution, it may be assumed that it could be effective.</p>	<i>Effectiveness of measures</i>	<i>Effective</i>	X	<i>Neutral</i>	<i>Ineffective</i>
<i>Effectiveness of measures</i>	<i>Effective</i>	X	<i>Neutral</i>	<i>Ineffective</i>		
<p>Effort required</p>	<p>This measure would need to be applied indefinitely.</p>					

e.g. period of time over which measure need to be applied to have results																						
Resources required ¹ e.g. cost, staff, equipment etc.	Efforts for a review of literature would be minimal, requiring only a short desk study by an expert. Efforts for surveillance would be dictated by findings from such a review, but would considered as low cost. This would require site visits to ascertain presence and delimit the area(s) where the species in (i) present and (ii) naturalized.																					
Side effects (incl. potential) – both positive and negative i.e. positive or negative side effects of the measure on public health, environment including non-targeted species, etc. For each of the side effect types please select one of the impact categories (with an 'X'), and provide a rationale, with supporting evidence and examples if possible.	<table border="1"> <tr> <td>Environmental effects</td> <td><i>Positive</i></td> <td></td> <td><i>Neutral or mixed</i></td> <td>X</td> <td><i>Negative</i></td> <td></td> </tr> <tr> <td>Social effects</td> <td><i>Positive</i></td> <td></td> <td><i>Neutral or mixed</i></td> <td>X</td> <td><i>Negative</i></td> <td></td> </tr> <tr> <td>Economic effects</td> <td><i>Positive</i></td> <td></td> <td><i>Neutral or mixed</i></td> <td>X</td> <td><i>Negative</i></td> <td></td> </tr> </table> <p><i>Rationale:</i> No side-effects are envisaged, but a potential positive side effect may result if/as surveys could also identify other alien invasive species.</p>	Environmental effects	<i>Positive</i>		<i>Neutral or mixed</i>	X	<i>Negative</i>		Social effects	<i>Positive</i>		<i>Neutral or mixed</i>	X	<i>Negative</i>		Economic effects	<i>Positive</i>		<i>Neutral or mixed</i>	X	<i>Negative</i>	
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Acceptability to stakeholders	<i>Acceptable</i>		<i>Neutral or mixed</i>		<i>Unacceptable</i>																	
Additional cost information ¹ When not already included above, or in the species Risk Assessment.	On-site surveys of high risk areas would require less effort than remote sensing plus required surveying (ground-truthing).																					
Level of confidence on the information provided ² Please select one of the confidence categories along with a statement to support the category chosen. See <i>Notes</i> section at the bottom of this document. NOTE – this is not related to the effectiveness of the measure	<table border="1"> <tr> <td><i>Inconclusive</i></td> <td></td> <td><i>Unresolved</i></td> <td></td> <td><i>Established but incomplete</i></td> <td>X</td> <td><i>Well established</i></td> <td></td> </tr> </table> <p><i>Rationale:</i> NPPOs could provide additional information, but it is assumed that there is confidence in this information.</p>	<i>Inconclusive</i>		<i>Unresolved</i>		<i>Established but incomplete</i>	X	<i>Well established</i>														
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Rapid eradication for new introductions - Measures to achieve eradication at an early stage of invasion, after an early detection of a new occurrence (cf. Article 17). This section assumes that the species is not currently present in a Member State, or part of a Member State's territory. **This table is repeated for each of the eradication measures identified.**

<p>Measure description Provide a description of the measure, and identify its objective</p>	<p>Manual eradication to cut all identified plants below ground level.</p> <p>Trees reproduce only by seeds, and no natural vegetative propagation has been reported (Pasiecznik et al., 2001). Hand clearance, or mechanical measures such as clearing/root ploughing using Caterpillar tractors, can be used to fell trees and uproot stumps. It is essential that roots are cut below ground level. Plants cut at or above ground level will otherwise coppice (i.e. resprout). For larger trees where removal of the root is considered difficult, stumps can be killed either by (i) burning of the stump, or (ii) application of a chemical stump treatment (see Chemical treatment). However, it is considered that chemical treatments would not be required in the EU, considering the restrictive size of invasions and restrictive size of plants present. Follow on treatments are also required to ensure that seedlings emerging from any seedbank are also removed.</p> <p>The only confirmed report of <i>Prosopis</i> in mainland Europe is in Almeria (two planted trees only: planted in 1988), south-eastern Spain (Pasiecznik and Peñalvo López, 2016). Elsewhere in the EU, it is reported as naturalized in a very limited area in the Canary Islands, Spain (Verloove, 2013, 2017). Here the species has been known since 2011 as an escape from cultivation in the drier southernmost parts of Gran Canaria. In 2015, it was recorded in several additional localities all in <i>barrancos</i> (seasonally dry valleys), and in one of these, the estuary of the barranco del Polvo in Arinaga, it is present in relative abundance and in various stages of development in natural coastal vegetation. Other reports of <i>P. juliflora</i> from Cyprus (Bovill, 1915; Frangos, 1923) and Italy in 1913 (Maniero, 2000) are considered invalid.</p>						
<p>Scale of application At what scale is the measure applied? What is the largest scale at which it has been successfully used? Please provide examples, with areas (km² or ha) if possible.</p>	<p>This method has been applied to small areas infestations in Australia, Djibouti, India, Kenya and Somaliland, amongst others, often only a few hectares, and not exceeding a few tens of hectares. The measure could be applied in all areas where <i>P. juliflora</i> is reported to be present (see above).</p>						
<p>Effectiveness of the measure</p>	<p><i>Effectiveness of measures</i></p>	<p><i>Effective</i></p>	<p>X</p>	<p><i>Neutral</i></p>		<p><i>Ineffective</i></p>	

<p>Is it effective in relation to its objective? Has the measure previously worked, failed? Please select one of the categories of effectiveness (with an 'X'), and provide a rationale, with supporting evidence and examples if possible.</p>	<p><i>Rationale:</i> It has been applied to the small infestations listed above, but follow-up actions are also needed over many years depending on the size of the soil seed bank. Without effective and long term follow up, eradication of small areas has not proved to be effective (e.g. Djibouti, India, Kenya).</p> <p>Due to the very restricted areas of <i>P. juliflora</i> reported in the EU at present, eradication is considered feasible. However, if it becomes established over large areas (which is not currently the case in the EU), there are no effective measures known to limit unintentional spread (see table above). To be certain that eradication can be undertaken at low cost, further information is required on the exact extent of <i>P. juliflora</i> population reported in Gran Canaria, as it is considered small but it could be larger. Additional surveys would be required to confirm these, alongside further literature reviews and surveys to assess the presence of any other populations.</p>						
<p>Effort required e.g. period of time over which measure need to be applied to have results</p>	<p>Eradication of the small areas where <i>P. juliflora</i> is present is estimated to be possible in only a short period (of days to a week) by a small work team. However, if left to become more significantly established, then more costly measures would need to be implemented, to reduce unintentional spread.</p> <p>Considering the limited naturalization in the EU, the areas should be monitored annually for at least five years, when a reassessment should be made. If no new seedlings have been reported and removed, monitoring can be reduced to every two years for at least 15 years, and then stop.</p>						
<p>Resources required¹ e.g. cost, staff, equipment etc.</p>	<p>Although very effective, manual clearing operations are labour-intensive and is practical only for small land holdings. If manual clearance is not undertaken immediately when areas are restricted and populations spread, then other methods may be required (see the Management section, below).</p>						
<p>Side effects (incl. potential) – both positive and negative i.e. positive or negative side effects of the measure on public health, environment including non-targeted species, etc. For each of the side effect types please select one of the impact categories (with an 'X'), and provide a rationale, with supporting evidence and examples if possible.</p>	<p>Environmental effects</p>	<p>Positive</p>	<input type="checkbox"/>	<p>Neutral or mixed</p>	<input checked="" type="checkbox"/>	<p>Negative</p>	<input type="checkbox"/>
	<p>Social effects</p>	<p>Positive</p>	<input type="checkbox"/>	<p>Neutral or mixed</p>	<input checked="" type="checkbox"/>	<p>Negative</p>	<input type="checkbox"/>
	<p>Economic effects</p>	<p>Positive</p>	<input type="checkbox"/>	<p>Neutral or mixed</p>	<input checked="" type="checkbox"/>	<p>Negative</p>	<input type="checkbox"/>
	<p><i>Rationale:</i> Manual cutting is by far the best method with limited areas/numbers of plants, being very targeted. If larger scale mechanical methods are used, this would likely also uproot other (non-target) species and is as such less desirable as a control method (and more costly).</p>						

<p>Acceptability to stakeholders e.g. impacted economic activities, animal welfare considerations, public perception, etc. Please select one of the categories of acceptability (with an 'X'), and provide a rationale, with supporting evidence and examples if possible.</p>	<p>Acceptability to stakeholders</p>	<p>Acceptable</p>	<p>X</p>	<p>Neutral or mixed</p>		<p>Unacceptable</p>		
<p>Additional cost information¹ When not already included above, or in the species Risk Assessment. - implementation cost - the cost of inaction - the cost-effectiveness - the socio-economic aspects</p>		<p>The cost of inaction, would increase significantly in the future as any management programme would have to take place on a larger scale and this would increase the cost of any measures.</p> <p>As there are very limited occurrences of <i>P. juliflora</i> in the EU in the natural environment, implementation costs for Member States would be relatively low.</p>						
<p>Level of confidence on the information provided² Please select one of the confidence categories along with a statement to support the category chosen. See <i>Notes</i> section at the bottom of this document.</p>	<p>Inconclusive</p>		<p>Unresolved</p>		<p>Established but incomplete</p>	<p>X</p>	<p>Well established</p>	
<p>Rationale: Manual eradication is known to be effective over small areas (Pasiiecznik et al., 2001), but there are no reports of complete eradication.</p>								

<p>Management - Measures to achieve management of the species once it has become widely spread within a Member State, or part of a Member State's territory. (cf. Article 19), i.e. not at an early stage of invasion (see Rapid eradication table above). These measures can be aimed at eradication, population control or containment of a population of the species. This table is repeated for each of the management measures identified.</p>	
<p>Measure description Provide a description of the measure, and identify its objective</p>	<p>Where widespread, prepare and implement national management plans</p> <p>As <i>P. juliflora</i> currently has only a very restricted distribution, it is proposed that only manual clearance will be necessary. Other tables follow, however, covering mechanical, chemical and biological control, but that will only be required if invasions become widespread, and should be implemented as part of a broader management plan).</p>

	<p>Where widespread, countries must prepare and implement eradication and containment/management plans. Management plans, or national strategies usually encompass a series of integrated measures depending upon the objective and costs. The individual measures include mechanical, chemical and biological control and are discussed separately in the following management tables, as approaches to manage <i>Prosopis</i> populations once it is widely established. The information in the following management tables is adapted from Pasiecznik et al. (2001) and CABI (2018), including those used on closely related <i>Prosopis</i> species, as it is considered that they can be applied equally on any species. Total tree kill may be possible with some treatments, but adequate techniques for preventing the re-introduction of seeds and re-establishment of trees have yet to be developed. It is considered that eradication over large areas is not possible using these techniques and, at best, only some form of integrated control is feasible.</p> <p>Due to the very restricted known current distribution in the EU, as explained in previous sections, it is considered that the only known populations (in Spain – Almeria and Gran Canaria) can be quickly eradicated at low cost, and risks of establishment are considered as low (EPPO, 2018). However, if this is not done in a timely manner, or further naturalizations are identified and found to be widespread, integrated management plans may need to be prepared and implemented, to control populations. These will include manual and mechanical techniques, alongside monitoring and surveillance to include early detection for countries most prone to risk, and ideally public awareness campaigns to prevent spread from existing populations or from botanic gardens in countries at high risk (as discussed above).</p>
<p>Scale of application At what scale is the measure applied? What is the largest scale at which it has been successfully used? Please provide examples, with areas (km² or ha) if possible.</p>	<p>These integrated action plans are usually developed at a national scale. The following is the known list of national strategies, resulting from a global review undertaken in 2013, during the author’s involvement in the preparation of a national strategy for Djibouti.</p> <p>Australia Australia was the first country to launch a national strategy on <i>Prosopis</i> in 2001 with the latest revision in 2012, and to date is the only country known to have such a strategy in place. Each state of Australia also has its own specific management plan for <i>Prosopis</i>. <i>Australian Weeds Committee, 2012. Mesquite (Prosopis spp.) strategic plan 2012–17. Weeds of National Significance, Australian Government Department of Agriculture, Fisheries and Forestry, Canberra, Australia. 37pp. www.weeds.org.au/WoNS/mesquite/docs/WEEDS-Mesquite-07-FINAL(18Mar13).pdf</i></p> <p>Ascension Island Significant attempts are known to have been made to control <i>Prosopis</i> on Ascension Island, and a management strategy is documented but no more recent information has been forthcoming.</p>

Belton, T, 2008. *Management strategy for Mexican thorn (Prosopis juliflora) on Ascension Island: An assessment of this species, and recommendations for management.* RSPB, UK, and the Ascension Island Government Conservation Department. 23pp.

Botswana

Recognising that the management and control of *Prosopis* is a transboundary issue, the Kalahari Namib Project, funded by UNEP GEF has supported the Government of Botswana to develop an Integrated National Mesquite (*Prosopis* species) Management (INMM) Strategy based on experiences from participating partners in Namibia and South Africa. Taken from the following release, - www.unccd.int/en/about-the-convention/the-bodies/the-cop/cop11/Pages/Side_Event_RegItemView.aspx?ItemID=56 (Sept 2013).

Eritrea

A draft framework of appropriate measures within a national action plan on *Prosopis* for Eritrea was prepared as part of a PhD thesis [Page 169-179], although no further information was elucidated about whether action has been taken based on this work.

Bokrezion H, 2008. *The ecological and socioeconomic role of Prosopis juliflora in Eritrea. An analytical assessment within the context of rural development in the Horn of Africa.* PhD Thesis, Johannes Gutenberg University, Mainz, Germany. 227pp.

Ethiopia

A national *Prosopis* management was prepared in 2002 resulting from a FAO consultancy mission.

Felker P, 2002. *Ethiopia - national plan for Prosopis.* FAO, Rome, Italy. 46pp.

However, this was never adopted and by 2008 no clear policy or strategy was in place - "At the national level there is no clear policy or strategy about control and management of Invasive Alien Species in general and *Prosopis* in particular (Anage et al., 2004; Fisehaye, 2006)", cited in:

Tegegn GG, 2008. *Experiences on Prosopis management. Case of Afar Region.* FARM Africa, Addis Ababa, Ethiopia. 35pp. [Page 11, also citing Anage et al., 2004].

Kenya

Kenya is the only country in Africa where a national *Prosopis* management strategy has been fully developed, where it awaiting final agreement before being submitted for approval. Contact skchoge@yahoo.com.

KFS/KEFRI, 2011. *Sectoral strategy for the management of prosopis species in Kenya, 2011–2015.* Kenya Forest Service (KFS) and Kenya Forestry Research Institute (KEFRI), and the Ministry of Forestry and Wildlife Development, Nairobi. 32pp.

Somalia

“In collaboration with the Ministry of Natural Resources, Somalia, Benadir Regional Assembly (Mogadishu Municipality) and UN-Habitat, CESVI and the Human Relief Foundation, work is currently underway in drafting a national urban strategy for the integrated management of *Prosopis* in urban areas of Somalia. It is anticipated that the national urban strategy will eventually form a component of a national *Prosopis* policy covering all geographical areas of Somalia as well as a range of eco-systems.”

Email received by Nick Pasiecznik on 28 November 2013 from: Dr. Andrew Adam-Bradford, Director - Horn of Africa Unit, Human Relief Foundation.

South Africa

South Africa developed 20-year vision as a *de facto* ‘management plan’, when over 50 stakeholders, representing all spheres of society and government, met in Kimberley in November 2001 to discuss the ‘status and long-term management of *Prosopis*’. The resulting declaration was: “In 20 years from now, invasive *prosopis* in Southern Africa will be under control and confined to areas where it can be managed to deliver sustainable benefits”. They envisaged, among others, development of new and value-adding utilisation programmes, and integrated agroforestry systems, including switching to benign varieties. Taken from:

Zimmermann H, Pasiecznik NM, 2005. Realistic approaches to the management of Prosopis species in South Africa. Policy brief. HDRA, Coventry, UK. 4pp.

www.gardenorganic.org.uk/pdfs/international_programme/SouthAfricaProsopisBrief.pdf

In addition, confirmation was received by Nick Pasiecznik (3 December 2013) in an email from Ross Shackleton of Stellenbosch University, South Africa, that he is in the process of producing a strategic plan for *Prosopis* management in South Africa, driven by Dave Richardson and Brian van Wilgen, published in 2017 (Shackleton et al., 2017).

Sudan

“During the early 1990s a popular opinion in parts of central Sudan and within the Sudanese Government had begun to consider *Prosopis* a noxious weed and a problematic tree species due to its aggressive ability to invade farmlands and pastures, especially in and around irrigated agricultural lands. As a consequence, *Prosopis* was deemed an invasive alien species, and on 26 February 1995, a presidential decree for its eradication [from everywhere in Sudan] was issued, which was followed by campaigning to execute the eradication.” Page 11, in: *Laxén J, 2007. Is prosopis a curse or a blessing? – An ecological-economic analysis of an invasive alien tree species in Sudan. Tropical Forestry Reports 32. VITRI, Helsinki, Finland. 203pp.*

<https://helda.helsinki.fi/bitstream/handle/10138/20611/isprosop.pdf?sequence=2>

However, it appears that no national strategy is in place as of 2012, as “the establishment of appropriate management plans of *Prosopis* is keenly demanded”, cited in the following paper.

	<p><i>Yoda K, Elbasit MA, Hoshino B, Nawata H, Yasuda H, 2012. Root system development of Prosopis seedlings under different soil moisture conditions. Journal of Arid Land Studies 22(1):13-16.</i></p> <p>[See also Standard PM3/67 'Guidelines for the management of invasive alien plants or potentially invasive alien plants which are intended for import or have been intentionally imported' (IPPC, 2010)].</p>																						
<p>Effectiveness of the measure Is it effective in relation to its objective? Has the measure previously worked, failed? Please select one of the categories of effectiveness (with an 'X'), and provide a rationale, with supporting evidence and examples if possible.</p>	<table border="1"> <tr> <td>Effectiveness of measures</td> <td><i>Effective</i></td> <td><input type="checkbox"/></td> <td><i>Neutral</i></td> <td><input checked="" type="checkbox"/></td> <td><i>Ineffective</i></td> <td><input type="checkbox"/></td> </tr> </table>	Effectiveness of measures	<i>Effective</i>	<input type="checkbox"/>	<i>Neutral</i>	<input checked="" type="checkbox"/>	<i>Ineffective</i>	<input type="checkbox"/>	<p><i>Rationale:</i> Significant efforts to contain <i>Prosopis</i> invasions have been implemented in Australia, but effectiveness has proved limited at best.</p>														
Effectiveness of measures	<i>Effective</i>	<input type="checkbox"/>	<i>Neutral</i>	<input checked="" type="checkbox"/>	<i>Ineffective</i>	<input type="checkbox"/>																	
<p>Effort required e.g. period of time over which measure need to be applied to have results</p>	See individual tables below for measure specific information.																						
<p>Resources required¹ e.g. cost, staff, equipment etc.</p>	No quantitative information is available. However, an initial estimate regarding the development of a national strategy could perhaps be assumed to cost less than €50,000, and possibly much less for an initial review. See individual tables below for measure specific information.																						
<p>Side effects (incl. potential) – both positive and negative i.e. positive or negative side effects of the measure on public health, environment including non-targeted species, etc. For each of the side effect types please select one of the impact categories (with an 'X'), and provide a rationale, with supporting evidence and examples if possible.</p>	<table border="1"> <tr> <td>Environmental effects</td> <td><i>Positive</i></td> <td><input type="checkbox"/></td> <td><i>Neutral or mixed</i></td> <td><input checked="" type="checkbox"/></td> <td><i>Negative</i></td> <td><input type="checkbox"/></td> </tr> <tr> <td>Social effects</td> <td><i>Positive</i></td> <td><input type="checkbox"/></td> <td><i>Neutral or mixed</i></td> <td><input checked="" type="checkbox"/></td> <td><i>Negative</i></td> <td><input type="checkbox"/></td> </tr> <tr> <td>Economic effects</td> <td><i>Positive</i></td> <td><input type="checkbox"/></td> <td><i>Neutral or mixed</i></td> <td><input checked="" type="checkbox"/></td> <td><i>Negative</i></td> <td><input type="checkbox"/></td> </tr> </table>	Environmental effects	<i>Positive</i>	<input type="checkbox"/>	<i>Neutral or mixed</i>	<input checked="" type="checkbox"/>	<i>Negative</i>	<input type="checkbox"/>	Social effects	<i>Positive</i>	<input type="checkbox"/>	<i>Neutral or mixed</i>	<input checked="" type="checkbox"/>	<i>Negative</i>	<input type="checkbox"/>	Economic effects	<i>Positive</i>	<input type="checkbox"/>	<i>Neutral or mixed</i>	<input checked="" type="checkbox"/>	<i>Negative</i>	<input type="checkbox"/>	<p><i>Rationale:</i> See individual tables below for measure specific information.</p>
Environmental effects	<i>Positive</i>	<input type="checkbox"/>	<i>Neutral or mixed</i>	<input checked="" type="checkbox"/>	<i>Negative</i>	<input type="checkbox"/>																	
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<p>Acceptability to stakeholders e.g. impacted economic activities, animal welfare considerations, public perception, etc.</p>	<table border="1"> <tr> <td>Acceptability to stakeholders</td> <td><i>Acceptable</i></td> <td><input checked="" type="checkbox"/></td> <td><i>Neutral or mixed</i></td> <td><input type="checkbox"/></td> <td><i>Unacceptable</i></td> <td><input type="checkbox"/></td> </tr> </table>	Acceptability to stakeholders	<i>Acceptable</i>	<input checked="" type="checkbox"/>	<i>Neutral or mixed</i>	<input type="checkbox"/>	<i>Unacceptable</i>	<input type="checkbox"/>	<p><i>Rationale:</i></p>														
Acceptability to stakeholders	<i>Acceptable</i>	<input checked="" type="checkbox"/>	<i>Neutral or mixed</i>	<input type="checkbox"/>	<i>Unacceptable</i>	<input type="checkbox"/>																	

<p>Please select one of the categories of acceptability (with an 'X'), and provide a rationale, with supporting evidence and examples if possible.</p>	<p>There can be no considered reason why any stakeholder would object to the management of known <i>P. juliflora</i> trees in the EU.</p>							
<p>Additional cost information ¹ When not already included above, or in the species Risk Assessment. - implementation cost - the cost of inaction - the cost-effectiveness - the socio-economic aspects Include quantitative &/or qualitative data, and case studies (incl. from countries outside the EU).</p>	<p>n/a</p>							
<p>Level of confidence on the information provided ² Please select one of the confidence categories along with a statement to support the category chosen. See <i>Notes</i> section at the bottom of this document.</p>	<p><i>Inconclusive</i></p>	<input type="checkbox"/>	<p><i>Unresolved</i></p>	<input type="checkbox"/>	<p><i>Established but incomplete</i></p>	<input type="checkbox"/>	<p><i>Well established</i></p>	<p>X</p>
<p><i>Rationale:</i> This information is based on significant previous knowledge of the author and numerous thorough reviews as cited and duplicated as required.</p>								

<p>Management - Measures to achieve management of the species once it has become widely spread within a Member State, or part of a Member State's territory. (cf. Article 19), i.e. not at an early stage of invasion (see Rapid eradication table above). These measures can be aimed at eradication, population control or containment of a population of the species. This table is repeated for each of the management measures identified.</p>	
<p>Measure description Provide a description of the measure, and identify its objective</p>	<p>Mechanical control. Mechanical site clearance involves tractor operations to remove trees, where roots are severed below ground level to ensure the tree is killed. These operations include root ploughing and chaining, which are often the most effective mechanical means. Root ploughing uses a mouldboard plough pulled behind a Caterpillar tractor, chaining involves pulling a heavy chain between two slow-moving Caterpillar tractors, with the effect of pulling over larger trees and uprooting them.</p>

<p>Scale of application At what scale is the measure applied? What is the largest scale at which it has been successfully used? Please provide examples, with areas (km² or ha) if possible.</p>	<p>On all areas where <i>P. juliflora</i> is identified as present.</p>						
<p>Effectiveness of the measure Is it effective in relation to its objective? Has the measure previously worked, failed? Please select one of the categories of effectiveness (with an 'X'), and provide a rationale, with supporting evidence and examples if possible.</p>	<p>Effectiveness of measures</p>	<p><i>Effective</i></p>		<p><i>Neutral</i></p>	<p>X</p>	<p><i>Ineffective</i></p>	
<p><i>Rationale:</i> For root ploughing, large trees must first be felled by hand, but this treatment has been used to remove stumps up to 50 cm in diameter without difficulty and has a treatment life of 20 years or more (Jacoby and Ansley, 1991). The soil should be neither too wet nor too dry for effective root ploughing. However, this method is one of the most expensive control treatments and is recommended only on deep soils that have a high potential for subsequent increased forage production (Jacoby & Ansley, 1991). For chaining, soil moisture is again important, with soil that is dry on the surface and moist below giving the optimal conditions. If the soil is too dry, the stem breaks leading to coppicing, if too wet, the soil and understorey are damaged (Jacoby & Ansley, 1991). Smaller, unbroken trees have to be removed by other means. Although expensive, this treatment is effective where there are many mature trees. It is most widely used following herbicide application to remove dead standing trees. Clearance with a biomass harvester produces wood chips that can be sold for energy production offsetting the operational costs (e.g. Felker, McLauchlan, Conkey & Brown, 1999).</p>							
<p>Effort required e.g. period of time over which measure need to be applied to have results</p>	<p>For root ploughing only a single pass is required, and leads to improved soil water conservation, and there is a chance to reseed with improved forage species. For chaining, a second pass in the opposite direction ensures that roots on all sides are severed to ease tree removal (Jacoby & Ansley, 1991).</p>						
<p>Resources required¹ e.g. cost, staff, equipment etc.</p>	<p>Heavy machinery (Caterpillar tractors) and specialist equipment (e.g root ploughs, chain, etc.), labour.</p>						
<p>Side effects (incl. potential) – both positive and negative i.e. positive or negative side effects of the measure on public health, environment including non-targeted species, etc. For each of the side effect types please select one of the impact</p>	<p>Environmental effects</p>	<p><i>Positive</i></p>		<p><i>Neutral or mixed</i></p>		<p><i>Negative</i></p>	<p>X</p>
<p><i>Rationale:</i> This a costly measure and would also destroy all other vegetation in the treated area.</p>							

categories (with an 'X'), and provide a rationale, with supporting evidence and examples if possible.								
Acceptability to stakeholders e.g. impacted economic activities, animal welfare considerations, public perception, etc. Please select one of the categories of acceptability (with an 'X'), and provide a rationale, with supporting evidence and examples if possible.	Acceptability to stakeholders	<i>Acceptable</i>		<i>Neutral or mixed</i>	X	<i>Unacceptable</i>		
	<i>Rationale:</i> Environmental concerns may make the implementation of this measure unacceptable, especially due to impacts on non-target species.							
Additional cost information ¹ When not already included above, or in the species Risk Assessment.	n/a							
Level of confidence on the information provided ² Please select one of the confidence categories along with a statement to support the category chosen. See <i>Notes</i> section at the bottom of this document.	<i>Inconclusive</i>		<i>Unresolved</i>		<i>Established but incomplete</i>	X	<i>Well established</i>	
	<i>Rationale:</i>							

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Measure description Provide a description of the measure, and identify its objective	Biological control <i>Prosopis</i> species continue to spread widely in parts of their native ranges where many native insect species (including bruchids, twig girdlers, psyllids and other injurious pests) are common components of the ecology. These regularly attack <i>Prosopis</i> but the trees have adapted to infestation by these pests and are still able to become invasive weeds over large tracts of land. But, several biological control programmes using species of seed-feeding bruchid beetles have been developed and implemented. The advantage with bruchids is their observed host specificity, with many species

	<p>found to feed only on <i>Prosopis</i>, and some only on a single species. Other insect species known to have a deleterious effect on native and exotic <i>Prosopis</i> in the Americas, mainly twig girdlers and psyllids, have also been suggested as possible biological control agents. The twig girdler <i>Oncideres limpida</i> attacks <i>P. pallida</i> in Brazil (Lima, 1994), whereas <i>Oncideres rhodostricta</i> is seen as a serious pest of <i>P. glandulosa</i> in the USA (Polk and Ueckert, 1973). Psyllids are known to severely affect the growth of <i>Prosopis</i> (Hodkinson, 1991) and have been suggested for use in controlling invasions.</p> <p>Most work on biological control of <i>Prosopis</i> to date has been carried out in South Africa and Australia, where several programmes are underway. The seed-feeding insects <i>Mimosetes protractus</i> and <i>Neltumius arizonensis</i> were introduced to eastern South Africa in conjunction with the bruchid beetles <i>Algarobius prosopis</i> and <i>A. bottimeri</i> for the control of invasive <i>Prosopis</i> species. <i>Neltumius arizonensis</i> and <i>A. prosopis</i> were successful in establishing themselves in large numbers and having a significant effect on <i>Prosopis</i> spp., whereas the other species were only found in low numbers (Hoffmann, Impson & Moran, 1993). Maximum damage to seeds occurred where grazing was controlled, as the multiplication and progress is hampered by livestock devouring pods before the insects destroy them. The same two bruchid species were also introduced to Ascension Island in an attempt to control <i>P. juliflora</i> which is present on 80% of the island, often in dense thickets. Two other species, one a psyllid and the other a mirid, were identified as attacking <i>P. juliflora</i> on Ascension Island and were thought to have been introduced accidentally from the Caribbean. The mirid <i>Rhinocloa</i> sp. causes widespread damage and is thought to lead to substantial mortality of trees (Fowler, 1998). In Australia, <i>Prosopis</i> infestations are at a relatively early stage and extreme care is being employed in the selection of suitable biological control agents, following the long history of problems caused there by plant and animal introductions. Insect species continue to be tested for their efficacy and host specificity as possible biological control agents of <i>Prosopis</i> species in Australia (e.g. van Klinken, 1999; van Klinken, Hoffmann, Zimmermann & Roberts, 2009). Besides the two <i>Algarobius</i> species, the sap-sucking psyllid <i>Prosopidosylla flava</i> and the leaf-tying moth, <i>Evippe</i> sp. have both been found to provide some control in Australia. Where identified as an invasive species in dry zone in northern Myanmar (e.g. Aung & Koike, 2015), there has been at least an initial focus on biological control agents for this forest invasive species (Than, 2011), with investigation for biological control agents conducted in Pyawbwe in January 2010. Damage was detected in the form of yellowing foliage and damage from pathogens around cuts during fuelwood harvesting, identified as <i>Fusarium</i> sp., <i>Tubercularia</i> sp. and <i>Nectria</i> sp., and small-scale trials have been initiated to examine the potential for these fungal pathogens to aid in biological control of <i>P. juliflora</i>.</p> <p>It should be borne in mind that the release of biological control agents is currently not regulated at EU level. Nevertheless national/regional laws are to be respected. Before any release of an alien species as a biological control agent an appropriate risk assessment should be made.</p>
Scale of application	Programs are usually initiated at country level, but it is best practice to engage with neighbouring countries who may also be potentially impacted by any resulting established bio-control agent.

<p>At what scale is the measure applied? What is the largest scale at which it has been successfully used? Please provide examples, with areas (km² or ha) if possible.</p>							
<p>Effectiveness of the measure Is it effective in relation to its objective? Has the measure previously worked, failed? Please select one of the categories of effectiveness (with an 'X'), and provide a rationale, with supporting evidence and examples if possible.</p>	<p>Effectiveness of measures</p>	<p><i>Effective</i></p>	<input type="checkbox"/>	<p><i>Neutral</i></p>	<input type="checkbox"/>	<p><i>Ineffective</i></p>	<p>X</p>
<p><i>Rationale:</i> Biological control of <i>Prosopis</i> spp. has been attempted in Australia and South Africa but has not proved effective (e.g. Rieks van Klinken, CSIRO, Australia, pers. comm.)</p>							
<p>Effort required e.g. period of time over which measure need to be applied to have results</p>	<p>Biocontrol agents require significant time (over many years) to undertake the research, but after release, the aim is to produce self-sustaining populations of the agent that will require no further effort.</p>						
<p>Resources required¹ e.g. cost, staff, equipment etc.</p>	<p>Significant prior research to identify the effects of any proposed biological control agent on non-target species in the affected area, and which is likely to run into, potentially, millions of Euros.</p>						
<p>Side effects (incl. potential) – both positive and negative i.e. positive or negative side effects of the measure on public health, environment including non-targeted species, etc. For each of the side effect types please select one of the impact categories (with an 'X'), and provide a rationale, with supporting evidence and examples if possible.</p>	<p>Environmental effects</p>	<p><i>Positive</i></p>	<input type="checkbox"/>	<p><i>Neutral or mixed</i></p>	<p>X</p>	<p><i>Negative</i></p>	<input type="checkbox"/>
<p>Social effects</p>							
<p><i>Positive</i></p>							
<p><i>Neutral or mixed</i></p>							
<p>X</p>							
<p><i>Negative</i></p>							
<p><input type="checkbox"/></p>							
<p>Economic effects</p>							
<p><i>Positive</i></p>							
<p><i>Neutral or mixed</i></p>							
<p>X</p>							
<p><i>Negative</i></p>							
<p><input type="checkbox"/></p>							
<p><i>Rationale:</i> Biological control has proved very effective in many instances, but there are instances where the agent spreads and causes impacts on native species (e.g. on <i>Opuntia</i> spp. in Central America).</p>							
<p>Acceptability to stakeholders e.g. impacted economic activities, animal welfare considerations, public perception, etc.</p>	<p>Acceptability to stakeholders</p>	<p><i>Acceptable</i></p>	<input type="checkbox"/>	<p><i>Neutral or mixed</i></p>	<p>x</p>	<p><i>Unacceptable</i></p>	<input type="checkbox"/>
<p><i>Rationale:</i></p>							

Please select one of the categories of acceptability (with an 'X'), and provide a rationale, with supporting evidence and examples if possible.	Introduction of further non-native species is likely to draw the interest and possible concern from the public and environmental lobby groups.								
Additional cost information ¹ When not already included above, or in the species Risk Assessment. - implementation cost - the cost of inaction - the cost-effectiveness - the socio-economic aspects Include quantitative &/or qualitative data, and case studies (incl. from countries outside the EU).	n/a								
Level of confidence on the information provided ² Please select one of the confidence categories along with a statement to support the category chosen. See <i>Notes</i> section at the bottom of this document.	<i>Inconclusive</i>		<i>Unresolved</i>		<i>Established but incomplete</i>	X	<i>Well established</i>		
<i>Rationale:</i>									

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Measure description Provide a description of the measure, and identify its objective	Chemical control <i>Note: This lists chemicals (PPP) that have been cited for use against the species. This does not mean the chemicals are available or legal to use and countries should check to ensure chemicals are licensed for use in their country. EU/national/local legislation on the use of plant protection products and biocides needs to be respected.</i>

	Chemical treatments involve the use of herbicides to kill trees, with the most effective being stem or aerial applications of systemic herbicides. The formulation and application of chemicals for trees of mixed ages and sizes within a stand is difficult. Although 2,4-D provided excellent suppression of top growth, few trees were actually killed and such chemical treatments had to be applied periodically to ensure that forage yields were maintained. Many herbicides and herbicide mixtures have been tested, mostly on <i>P. glandulosa</i> . Potential environmental damage from widespread use of herbicides must also be taken into consideration.						
Scale of application At what scale is the measure applied? What is the largest scale at which it has been successfully used? Please provide examples, with areas (km ² or ha) if possible.	On all affected areas in the endangered area.						
Effectiveness of the measure Is it effective in relation to its objective? Has the measure previously worked, failed? Please select one of the categories of effectiveness (with an 'X'), and provide a rationale, with supporting evidence and examples if possible.	Effectiveness of measures	<i>Effective</i>	X	<i>Neutral</i>		<i>Ineffective</i>	X
	<p><i>Rationale:</i></p> <p>Effectiveness is dependent upon chemical uptake, which in <i>Prosopis</i> is limited by the thick bark, woody stems and small leaves with a protective waxy outer layer. The most effective chemical in the USA is clopyralid, but dicamba, picloram and triclopyr have also been successfully used, either alone or in combination (Jacoby & Ansley, 1991). In India, ammonium sulfamate was successful in killing <i>P. juliflora</i> trees and as a stump treatment (Panchal & Shetty, 1977). Use of chemical alone have proved ineffective in control large areas of prosopis invasions, such as in the USA on <i>P. glandulosa</i> (see Pasiecznik et al., 2001). However, in restricted areas, chemical have proved effective, also as cut stump treatments.</p>						
Effort required e.g. period of time over which measure need to be applied to have results	The effort required is significant, either as (relatively ineffective) aerial treatments, basal bark applications or treatment of cut stumps. Infested sites often needed spraying every 5-7 years.						
Resources required ¹ e.g. cost, staff, equipment etc.	Trained labour, chemicals (which can be costly), application equipment (see below).						
Side effects (incl. potential) – both positive and negative i.e. positive or negative side effects of the measure on public health, environment including non-targeted species, etc.	Environmental effects	<i>Positive</i>		<i>Neutral or mixed</i>		<i>Negative</i>	X
	Social effects	<i>Positive</i>		<i>Neutral or mixed</i>	X	<i>Negative</i>	
	Economic effects	<i>Positive</i>		<i>Neutral or mixed</i>		<i>Negative</i>	X
	<i>Rationale:</i>						

<p>For each of the side effect types please select one of the impact categories (with an 'X'), and provide a rationale, with supporting evidence and examples if possible.</p>	<p>The use of chemicals must have negative environmental effects from chemical residues, especially from foliar applications, less so for basal bark or cut stump treatments, yet these cannot be discounted. Impacts on human health of the applications may also be possible unless all safety precautions are strictly adhered to.</p>							
<p>Acceptability to stakeholders e.g. impacted economic activities, animal welfare considerations, public perception, etc. Please select one of the categories of acceptability (with an 'X'), and provide a rationale, with supporting evidence and examples if possible.</p>	<p>Acceptability to stakeholders</p>	<p><i>Acceptable</i></p>	<input type="checkbox"/>	<p><i>Neutral or mixed</i></p>	<input type="checkbox"/>	<p><i>Unacceptable</i></p>	<p>X</p>	
<p><i>Rationale:</i> There may be objections to the use of chemicals, especially in natural areas or regional parks, and especially where other safer measures (such as manual or mechanical removal) are available and effective.</p>								
<p>Additional cost information¹ When not already included above, or in the species Risk Assessment.</p>	<p>Information from the USA has indicated a high cost of chemical control (e.g. Jacoby & Ansley, 1991)</p>							
<p>Level of confidence on the information provided² Please select one of the confidence categories along with a statement to support the category chosen. See <i>Notes</i> section at the bottom of this document.</p>	<p><i>Inconclusive</i></p>	<input type="checkbox"/>	<p><i>Unresolved</i></p>	<input type="checkbox"/>	<p><i>Established but incomplete</i></p>	<p>X</p>	<p><i>Well established</i></p>	<input type="checkbox"/>
<p><i>Rationale:</i> Further work is required to establish the effectiveness of chemical control.</p>								

Bibliography

This is taken from the Pest Risk Analysis for *Prosopis juliflora* undertaken for EU/EPPO region as part of the LIFE project in 2017 (EPPO, 2018). And whereas not all are cited in the preceding document, they are considered to provide a comprehensive and current bibliography.

Ahmad, R., Ismail, S. & Khan, D. (1996). Use of *Prosopis* in Arab/Gulf states, including possible cultivation with saline water in deserts. In: P. Felker & J. Moss (Eds). *Prosopis: Semiarid Fuelwood and Forage Tree; Building Consensus for the Disenfranchised*. Kingsville, Texas, USA: Center for Semi-Arid Resources, 1.41-1.52.

Andreu, J. & Vilà, M. (2010). Risk analysis of potential invasive plants in Spain. *Journal for Nature Conservation*, 18(1):34–44.

Archer, S. (1995). Tree-grass dynamics in a *Prosopis*-thornscrub savanna parkland -reconstructing the past and predicting the future. *Ecoscience*, 2:83-99.

Aung, T. & Koike, F. (2015). Identification of invasion status using a habitat invasibility assessment model: The case of *Prosopis* species in the dry zone of Myanmar. *Journal of Arid Environments*, 120:87–94.

Ayanu, J., Jentsch, A., Müller-Mahn, D., Rettberg, S., Romankiewicz, C. & Koellner, T. (2015). Ecosystem engineer unleashed: *Prosopis juliflora* threatening ecosystem services? *Regional Environmental Change*, 15(1):155–167.

Benson, L. 1941. The mesquites and screwbeans of the United States. *American Journal of Botany*, 28:748-754.

Bentham, G. (1875). Revision of the suborder Mimoseae. *The Transactions of the Linnean Society of London*, 30:335-664.

Bovill, A.K. (1915). *Report on Plantation Works in Cyprus 1879 to 1914*. Nicosia: Government Printing Office.

Burkart, A. (1940). Materiales para una monografia del genero *Prosopis* (Leguminosae). *Darwiniana*, 4:57-128.

Burkart, A. (1976). A monograph of the genus *Prosopis* (Leguminosae, subfam. Mimosoideae). (Part 1 and 2) Catalogue of the recognized species of *Prosopis*. *Journal of the Arnold Arboretum*, 57:217-246, 450-525.

CABI. (2018). *Prosopis juliflora*. In: *Invasive Species Compendium*. CABI, Wallingford, UK.

Catalano, S.A., Vilardi, J.C., Tosto, D. & Saidman, B.O. (2008). Molecular phylogeny and diversification history of *Prosopis* (Fabaceae: Mimosoideae). *Biological Journal of the Linnean Society*, 93:621-640.

- Chaudhry, B. & Vijayaraghavan, M.R. (1992). Structure and function of the anther gland in *Prosopis juliflora* (leguminosae, mimosoideae) - a histochemical analysis. *Phyton-Annales Rei Botanicae*, 32:1-7.
- Chinnimani, S. (1998). Ecology of succession of *Prosopis juliflora* in the ravines of India. In: J.C. Tewari, N.M. Pasiecznik, L.N. Harsh & P.J.C Harris (Eds.) *Prosopis Species in the Arid and Semi-Arid Zones of India*. Coventry, UK: Prosopis Society of India and the Henry Doubleday Research Association, 21-22.
- Choge, S.K., Ngujiri, F.D., Kuria, M.N., Busaka, E.A. & Muthondeki, J.K. (2002). The status and impact of *Prosopis* spp. in Kenya. KEFRI, Nairobi, Kenya.
- Choge, S., Clement, N., Gitonga, M. & Okuye, J. (2012). Status report on commercialization of prosopis tree resources in Kenya. Technical report for KEFRI/KFS Technical Forest Management and Research Liaison Committee. KEFRI, Nairobi, Kenya.
- Csurhes, S.M. (1996). Pest Status Review Series - Land Protection Branch: Mesquite (*Prosopis* spp.) in Queensland. Queensland, Australia: Department of Natural Resources.
- D'Antoni, H.L. & Solbrig, O.T. (1977). Algarrobos in South American cultures: past and present. In: B.B. Simpson (Ed.) *Mesquite: Its Biology in Two Desert Ecosystems*. Stroudsburg, Pennsylvania, USA: Dowden, Hutchinson and Ross, 189-200.
- Dhyani, A., Singh, B.P., Arora, N., Jain, V.K. & Sridhara, S. (2008). A clinically relevant major cross-reactive allergen from mesquite tree pollen. *European Journal of Clinical Investigation*, 38(10):774-781.
- Diagne, O. (1992). Current developments on *Prosopis* species in Senegal including work on nitrogen fixation. Dutton RW, ed. *Prosopis Species: Aspects of their Value, Research and Development*. University of Durham, UK: CORD, 47-60.
- Dufour-Dror, J.M. & Shmida, A. (2017). Invasion of alien *Prosopis* species in Israel, the West Bank and western Jordan: characteristics, distribution and control perspectives. *BioInvasions Records*, 6(1):1-7.
- El Fadl, M.A. (1997). Management of *Prosopis juliflora* for use in agroforestry systems in the Sudan. *Tropical Forestry Reports* 16. Helsinki, Finland: University of Helsinki.
- El-Keblawy, A. & Al-Rawai, A. (2007). Impacts of the invasive exotic *Prosopis juliflora* (Sw.) D.C. on the native flora and soils of the UAE. *Plant Ecology*, 190(1):23-35.
- EPPO. (2018). *Pest risk analysis for Prosopis juliflora*. EPPO, Paris.
- Esbenshade, H.W. (1980). Kiawe: a tree crop in Hawaii. *International Tree Crops Journal*, 1:125-130.

- Fonseca, N.C., Albuquerque, A.S., Leite, M.J. de H. & Lira, C.S. de. (2016). Similarity floristic and biological colonization of *Prosopis juliflora* [(Sw) DC] along the Paraíba river. *Nativa: Pesquisas Agrárias e Ambientais*, 4(6):392-397.
- Felker, P. (1979). Mesquite an all purpose leguminous arid land tree. In: G.A Ritchie (Ed.), *New Agricultural Crops*. American Association Advance. Sci. Publ., Bowder, USA, 38:89-125.
- Felker, P., Meyer, J.M. & Gronski, S.J. (1990). Application of self-thinning in mesquite (*Prosopis glandulosa* var *glandulosa*) to range management and lumber production. *Forest Ecology and Management*, 31:225-232.
- Felker, P., Clark, P.R., Nash, P., Osborn, J.F. & Cannell, G.H. (1982). Screening *Prosopis* (mesquite) for cold tolerance. *Forest Science*, 28:556-562.
- Felker, P., McLauchlan, R.A., Conkey, A. & Brown, S. (1999). Case study: development of a swath harvester for small diameter (<10 cm) woody vegetation. *Biomass and Bioenergy*, 17:1-17.
- Fowler, S.V. (1998). *Report on the invasion, impact and control of 'Mexican thorn', Prosopis juliflora, on Ascension Island*. Ascot, UK: CABI Bioscience.
- Frangos, G. (1923). Exotic trees now existing in Cyprus. *The Cyprus Agricultural Journal*, 18: 86-89.
- Goel, V.L. & Behl, H.M. (1996). Phenological traits of some woody species grown on stress soil. *Journal of the Indian Botanical Society*, 75:11-16.
- Habit, M.A. & Saavedra, J.C. (Eds.). (1990). *The Current State of Knowledge on Prosopis juliflora*. II International Conference on Prosopis, Recife, Brazil, 25-29 August, 1986. Rome, Italy: FAO.
- Harding, G.B. (1991). Sheep can reduce seed recruitment of invasive *Prosopis* species. *Applied Plant Science*, 5(1):25-27.
- Harris, P.J.C., Pasiecznik, N.M., Smith, S.J., Billington, J. & Ramirez, L. (2003). Differentiation of *Prosopis juliflora* (Sw.) DC. and *P. pallida* (H. & B. ex. Willd.) H.B.K. using foliar characters and ploidy. *Forest Ecology and Management*, 180:153-164.
- Hodkinson, I.D. (1991). New World legume-feeding psyllids of the genus *Aphalaroida* Crawford (Insecta: Homoptera: Psylloidea). *Journal of Natural History*, 25(5):1281-1296.
- Hoffmann, J.H., Impson, F.A.C. & Moran, V.C. (1993). Competitive interactions between two bruchid species (*Algarobius* spp.) introduced into South Africa for biological control of mesquite weeds (*Prosopis* spp.). *Biological Control*, 3(3):215-220.

- Hyde, E.A., Pasiecznik, N. & Harris, P.J.C. (1990). Evaluation of multi-purpose trees in southern Spain. *Nitrogen Fixing Tree Research Reports*, 8:73-74.
- IPPC. (2010). *Glossary of phytosanitary terms*. International Standards for Phytosanitary Measures ISPM) No. 5. Secretariat of the International Plant Protection Convention, FAO, Rome.
- Jacoby, P. & Ansley, R.J. (1991). Mesquite: classification, distribution, ecology and control. In: L.F. James, J.O. Evans, M.H. Ralphs, R.D. Child (Eds). *Noxious Range Weeds*. Boulder, Colorado, USA: Westview Press.
- Janssen, J.A.M., et al. (2016). *European Red List of habitats. Part 2. Terrestrial and freshwater habitats*. European Union, Brussels, Belgium. 38pp.
- Johnston, M.C. (1962). The North American Mesquites, *Prosopis* Sect. Algarobia (Leguminosae). *Brittonia*, 14:72-90.
- Kaur, R., Gonzáles, W.L., Llambi, L.D., Soriano, P.J., Callaway, R.M., Rout, M.E., Gallaher, T.J & Inderjit. (2012). Community impacts of *Prosopis juliflora* invasion: biogeographic and congeneric comparisons. *PLoS ONE*, 7(9):e44966.
- Khan, A.A. (1961). Efficacy of "Fernozone" to prevent resprouting of cut-back mesquite. *Pakistan Journal of Forestry*, 11:375-377.
- Khan, D., Ahmad, R. & Ismail, S. (1987). Germination, growth and ion regulation in *Prosopis juliflora* (Swartz) DC under saline conditions. *Pakistan Journal of Botany*, 19:131-138.
- Killian, S. & McMichael, J. (2004). The human allergens of mesquite (*Prosopis juliflora*). *Clinical and Molecular Allergy*, 2(8).
- Kool, M., van Steenberg, F., Haile, A.M., Mohamed, Y.A. & Nzumira, H. (2014). Controlling and/or using *Prosopis juliflora* in spate irrigation systems. Practical Note 25. Spate Irrigation Network.
- Kumar, S. & Mathur, M. (2014). Impact of invasion by *Prosopis juliflora* on plant communities in arid grazing lands. *Tropical Ecology*, 55(1):33-46.
- Landeras, G., Alfonso, M., Pasiecznik, N.M., Harris, P.J.C & Ramírez, L. (2006). Identification of *Prosopis juliflora* (Sw.) DC. and *P. pallida* (H. & B. ex. Willd.) H.B.K. using molecular markers. *Biodiversity and Conservation*, 15:1829-1844.
- Lima, P.C.F. (1994). *Comportamento silvicultural de especies de Prosopis, em Petrolina-PE, Regio Semi-Arida Brasileira*. PhD Thesis. Curitiba, Brazil: Universidade Federal do Parana.
- Little, E.L. & Wadsworth, F.H. (1964). *Common trees of Puerto Rico and the Virgin Islands*. US Department of Agriculture, Agricultural Handbook 249.

- Lowe, S., Browne, M., Boudjelas, S. & De Poorter, M. (2000). *100 of the World's Worst Invasive Alien Species A selection from the Global Invasive Species Database*. Invasive Species Specialist Group (ISSG), the World Conservation Union (IUCN), Gland, Switzerland. 12pp.
- Luna, R.K. (1996). *Prosopis juliflora* (Swartz) DC. In: *Plantation Trees*. Delhi, India: International Book Distributors.
- Maniero, F. (2000). *Fitocronologia d'Italia*, 290 pp. Casa Editrice Leo S. Olschki, Firenze (Italy).
- Maundu, P., Kibet, S., Morimoto, Y., Imbumi, M. & Adekar, R. (2009). Impact of *Prosopis juliflora* on Kenya's semi-arid and arid ecosystems and local livelihoods. *Biodiversity*, 10(2-3):33.
- Meikle, R.D. (1977). *Flora of Cyprus*. Vol. 1. London: Bentham-Moxon Trust, Royal Botanic Gardens, Kew.
- Meroni, M., Ng, W. T., Rembold, F., Leonardi, U., Atzberger, C., Gadain, H., & Shaiye, M. (2017). Mapping *Prosopis juliflora* in west Somaliland with Landsat 8 satellite imagery and ground information. *Land Degradation & Development*, 28(2), 494-506.
- Mohan, N.P. (1940). The Mesquite. *Punjab Forestry Record (Silva Publ.)* 9.
- Mutha, N. & Burman, U. (1998). Effect of seed weight and sowing depth on germination and seedling quality of *Prosopis juliflora*. In: J.C. Tewari, N.M. Pasiecznik, L.N. Harsh, P.J.C Harris (Eds). *Prosopis Species in the Arid and Semi-Arid Zones of India*. Coventry, UK: *Prosopis Society of India and the Henry Doubleday Research Association*, 43-45.
- Muthana, K.D. (1974). A note on frost susceptibility in arid zone trees. *Annals of Arid Zone* 13:370-373.
- Mwangi, E. & Swallow, B. (2005). *Invasion of Prosopis juliflora and local livelihoods: Case study from the Lake Baringo area of Kenya*. Nairobi, Kenya: World Agroforestry Centre, 68 pp.
- Ntshidi, Z., Dzikiti, S., Mazvimavi, D., Bugan, R.D.H., Le Maitre, D.C., Gush, M.B. & Jovanovic, N.Z. (2015). Comparative use of groundwater by invasive alien *Prosopis* spp and co-occurring indigenous v. karroo in a semi-arid catchment. In: 14th Biennial Ground Water Division Conference and Exhibition, 21-23 September 2015, Muldersdrift, South Africa, p1-10.
- Palacios, R.A. (2006). Mezquites Mexicanos: biodiversidad y distribución geográfica. *Bol. Sociedad Argentina Botanica*, 41(1-2):91-121.
- Palacios, R.A., Burghardt, A.D., Frías-Hernández, J.T., Olalde-Portugal, V., Grados, N., & Martínez Vega, O. de la. (2012). Comparative study (AFLP and morphology) of three species of *Prosopis* of the section *Algarobia*: *P. juliflora*, *P. pallida*, and *P. limensis*. Evidence for resolution of the "*P. pallida*-*P. juliflora* complex". *Plant Systematics and Evolution*, 298:165-171.

Panchal, Y.C. & Shetty, P. (1977). *Chemical control of Prosopis juliflora* (Sw) DC. Program and Abstracts of Papers, Weed Science Conference and Workshop in India, No. 153.

Panetta, F.D. & Carstairs, S.A. (1989). Isozymic discrimination of tropical Australian populations of mesquite (*Prosopis* spp.): implications for biological control. *Weed Research*, UK, 29(3):157-165

Pasha, S.V., Satish, K.V., Reddy, C.S., Rao, P.V.V.P. & Jha, C.S. (2014). Satellite image based quantification of invasion and patch dynamics of mesquite (*Prosopis juliflora*) in Great Rann of Kachchh, Kachchh biosphere reserve, Gujarat, India. *Journal of Earth System Science*, 123(7):1481-1490.

Pasiecznik, N.M. (1998). *Prosopis* and provenance research in Cape Verde. In: J.C. Tewari, N.M. Pasiecznik, L.N. Harsh, P.J.C Harris (Eds). *Prosopis Species in the Arid and Semi-Arid Zones of India*, pp73-76. Proceedings of a Conference, 21-23 November 1993. CAZRI, Jodhpur, India. The *Prosopis* Society of India and the Henry Doubleday Research Association, Coventry, UK.

Pasiecznik, N., Choge, S., Fre, Z., Tsegay, B. & Parra, F. (2015). *The Great Green Forest is here and expanding all on its own: A call for action*. International Conference on Resilience, Research and Innovation, Djibouti, 26-28 October 2015.

Pasiecznik, N.M. (2013). *Preliminary survey of prosopis invasions in Djibouti*. TCP/DJI/3303(A). Report of a mission, 21 April – 13 May 2013. FAO, Djibouti. 16pp.

Pasiecznik, N.M., Choge, S.K., Trenchard, E.J. & Harris, P.J.C. (2012). Improving food security in famine-prone areas using invasive and underutilised prosopis trees. *Food Chain* 2(2): 197-206.

Pasiecznik, N.M., Felker, P., Harris, P.J.C., Harsh, L.N., Cruz, G., Tewari, J.C., Cadoret, K. & Maldonado, L.J. (2001). *The Prosopis juliflora - Prosopis pallida Complex: A Monograph*. HDRA, Coventry, UK. 162pp.

Pasiecznik, N.M., Harris, P.J.C. & Smith, S.J. (2004). *Identifying Tropical Prosopis Species: A Field Guide*. HDRA, Coventry, UK. 30pp.

Pasiecznik, N.M., Ould Mohamed Ali, A., Nourissier-Mountou, S., Danthu, P., Murch, J., McHugh, M.J. & Harris, P.J.C. (2006). Discovery of a life history shift: precocious flowering in an introduced population of *Prosopis*. *Biological Invasions*, 8:1681-1687.

Pasiecznik, N.M. & Weerawardane, N.D.R. (2011). Invasion of the exotic tree species *Prosopis juliflora* Sw. (DC) in Puttalam District – its spread, new records, and the need for action. *Sri Lankan Forester*, 32/33: 43-50.

- Pasiecznik, N.M. & Peñalvo López, E. (2016). 25 year results of a dryland tree trial, and an invasive risk assessment of introduced species [Resultados tras 25 años de un ensayo de reforestación de zonas semiáridas, y el riesgo de introducción de especies exóticas]. *Zonas Áridas*, 16(1): 52-71.
- Perry, G. (1998). *Prosopis*. *Flora of Australia*, 12:7-13.
- Polk, K.L. & Ueckert, D.N. (1973). Biology and ecology of a mesquite twig girdler, *Oncideres rhodosticta*, in West Texas. *Annals of the Entomological Society of America*, 66(2):411-417.
- Raizada, M.B. & Chatterji, R.N. (1954). A diagnostic key to the various forms of introduced mesquite (*Prosopis juliflora* DC). *Indian Forester*, 80:675-680.
- Reddy, C.V.K. (1978). *Prosopis juliflora*, the precocious child of the plant world. *Indian Forester*, 104:14-18.
- Sareen, T.S. & Yadav, S.K. (1987). Self-incompatibility in *Prosopis juliflora* DC (Leguminosae). *Incompatibility Newsletter*, 19:66-68.
- Shackleton, R.T., Le Maitre, D.C., Pasiecznik, N.M. & Richardson, D.M. (2014). *Prosopis*: A global assessment of the biogeography, benefits, impacts and management of one of the world's worst woody invasive plant taxa. *AoB PLANTS*. doi: 10.1093/aobpla/plu027
- Shackleton, R.T., Le Maitre, D.C., van Wilgen, B.W. & Richardson, D.M. (2017). Towards a national strategy to optimise the management of a widespread invasive tree (*Prosopis* species; mesquite) in South Africa. *Ecosystem Services*, 27:.242-252.
- Sherry, M., Smith, S., Patel, A., Harris, P., Hand, P., Trenchard, L. & Henderson, J. (2011). RAPD and microsatellite transferability studies in selected species of *Prosopis* (section *Algarobia*) with emphasis on *Prosopis juliflora* and *P. pallida*. *Journal of Genetics*, 90:251-264.
- Simpson, B.B. (Ed.). (1977). *Mesquite, its Biology in Two Desert Shrub Ecosystems*. Stroudsburg, Pennsylvania, USA: Dowden, Hutchinson & Ross.
- Sinha, A.K., Shirke, P.A., Pathre, U. & Sane, P.V. (1997). Sucrose-phosphate synthase in tree species: light/dark regulation involves a component of protein turnover in *Prosopis juliflora* (SW DC). *Biochemistry and Molecular Biology International*, 43:421-431.
- Solbrig, O. & Cantino, P. (1975). Reproductive adaptations in *Prosopis* (Leguminosae-Mimosoideae). *Journal of the Arnold Arboretum*, 56(2):185-210.
- Srinivasu, V. & Toky, O. (1996). Effect of alkalinities on seed germination and seedling growth of important arid trees. *Indian Journal of Forestry*, 19:227-233.
- Sukhorukov, A.P., Verloove, F., Ángeles Alonso, M., Belyaeva, I.V., Chapanoe, C., Crespo, M.B., El Aouni, M.H., El Moknif, R., Maroyi, A., Shekede, M.D., Vicente, A., Dreyer, A. & Kushunina, M. (2017). Chorological and taxonomic notes on African plants, 2. *Botany Letters*, <http://dx.doi.org/10.1080/23818107.2017.1311281>

- Tabosa, I.M., Riet-Correa, F., Barros, S.S., Summers, B.A., Simões, S.V., Medeiros, R.M. & Nobre, V.M. (2006). Neurohistologic and ultrastructural lesions in cattle experimentally intoxicated with the plant *Prosopis juliflora*. *Veterinary Pathology*, 43(5):695-701.
- Tavares, J. de P. & Barros, D. (2016). Reforestation as a strategy in the restoration of shallow soils and recharge of groundwater in Cabo Verde. *Nature and Fauna*, 30(2):56-61.
- Tewari, J.C., Harris, P.J.C., Harsh, L.N., Cadoret, K. & Paseicznik, N.M. (2001). *Managing Prosopis juliflora (Vilayati babul) - a technical manual*. 96pp. CAZRI, HDRA. Jodhpur, India, Coventry, UK. Central Arid Zone Research Institute (CAZRI), Jodhpur, India and Henry Doubleday Research Association (HDRA), Ryton on Dunsmore, UK. ISBN:0-905343-27-1.
- Than, W.W. (2011). *An initial focus on biological control agents for the forest invasive species Prosopis juliflora in the dry zone of Myanmar*. In: XIII International Symposium on Biological Control of Weeds, Session 4, Target and Agent Selection. p178.
- Timyan, J. (1996). *Bwa yo: important trees of Haiti*, 418 pp.
- The Plant List. (2017). The Plant List Version 1. UK: Royal Botanic Gardens, Kew and Missouri Botanical Gardens. <http://www.theplantlist.org/>
- Trenchard, E.J., Harris, P.J.C., Smith, S.J. & Pasiecznik, N.M. (2008). A review of ploidy in the genus *Prosopis* (Leguminosae). *Biological Journal of the Linnean Society*, 156:425-438.
- USDA-ARS. (2017). *Germplasm Resources Information Network (GRIN)*. Online Database. National Germplasm Resources Laboratory, Beltsville, USA.
- Van Klinken, R.D. (1999). Developmental host-specificity of *Mozena obtusa* (Heteroptera: Coreidae), a potential biocontrol agent for *Prosopis* species (mesquite) in Australia. *Biological Control*, 16(3):283-290.
- Van Klinken, R.D., Hoffmann, J.H., Zimmermann, H.G. & Roberts, A.P. (2009). *Prosopis* spp. (Leguminosae). In: R. Muniappan, G.V.P Reddy & A. Raman (Eds.) *Biological Control of Tropical Weeds Using Arthropods*. Cambridge, UK, 353-377.
- Verloove, F. (2013). New xenophytes from Gran Canaria (Canary Islands, Spain), with emphasis on naturalized and (potentially) invasive species. *Collectanea Botanica*, 32:59–82
- Verloove, F. (2017). New xenophytes from the Canary Islands (Gran Canaria and Tenerife; Spain). *Acta Bot. Croat.* 76 (2), 120–131
- Wakie, T.T., Hoag, D., Evangelista, P.H., Luizza, M. & Laituri, M. (2016). Is control through utilization a cost effective *Prosopis juliflora* management strategy? *Journal of Environmental Management*, 168:74-86.

Warrag, M.O.A. (1994). Autotoxicity of mesquite (*Prosopis juliflora*) pericarps on seed germination and seedling growth. *Journal of Arid Environments*, 27:79-84.

Weber, E. (2003). *Invasive plant species of the world: a reference guide to environmental weeds*. CABI Publishing, Wallingford, UK. 548pp.

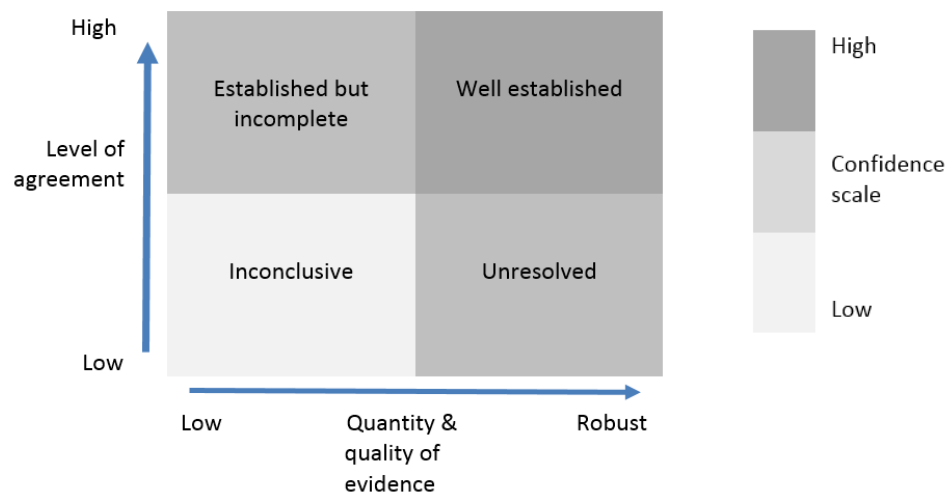
Wise, R.M., van Wilgen, B.W. & Le Maitreb, D.C. (2012). Costs, benefits and management options for an invasive alien tree species: The case of mesquite in the Northern Cape, South Africa. *Journal of Arid Environments*, 84:80-90.

Zachariades, C., Hoffmann, J.C. & Roberts, A.P. (2011). Biological control of mesquite (*Prosopis* species) (*Fabaceae*) in South Africa. *African Entomology*, 19:402-415.

Notes

1. Costs information. The assessment of the potential costs shall describe those costs quantitatively and/or qualitatively depending on what information is available. This can include case studies from across the Union or third countries.

2. Level of confidence¹: based on the quantity, quality and level of agreement in the evidence.



- **Well established:** comprehensive meta-analysis² or other synthesis or multiple independent studies that agree.
- **Established but incomplete:** general agreement although only a limited number of studies exist but no comprehensive synthesis and, or the studies that exist imprecisely address the question.
- **Unresolved:** multiple independent studies exist but conclusions do not agree.
- **Inconclusive:** limited evidence, recognising major knowledge gaps

3. Citations and bibliography. The APA formatting style for citing references in the text and in the bibliography is used.

e.g. Peer review papers will be written as follows:

In text citation: (Author & Author, Year)

In bibliography: Author, A. A., & Author, B. B. (Publication Year). Article title. *Periodical Title*, Volume(Issue), pp.-pp.

(see <http://www.waikato.ac.nz/library/study/referencing/styles/apa>)

¹ Assessment of confidence methodology is taken from IPBES. 2016. Guide on the production and integration of assessments from and across all scales (IPBES-4-INF-9), which is adapted from Moss and Schneider (2000).

² A statistical method for combining results from different studies which aims to identify patterns among study results, sources of disagreement among those results, or other relationships that may come to light in the context of multiple studies.